

# Effects of Urbanization, Soil Property and Vegetation Configuration on Soil Infiltration of Urban Forest in Changchun, Northeast China

WANG Peijiang<sup>1,2</sup>, ZHENG Haifeng<sup>1</sup>, REN Zhibin<sup>1</sup>, ZHANG Dan<sup>1</sup>, ZHAI Chang<sup>1,2</sup>, MAO Zhixia<sup>1,2</sup>, TANG Ze<sup>1,2</sup>, HE Xingyuan<sup>1,2</sup>

(1. Key Laboratory of Wetland Ecology and Environment, Northeast Institute of Geography and Agroecology, Chinese Academy of Sciences, Changchun 130102, China, 2. University of Chinese Academy of Sciences, Beijing 100049, China)

**Abstract:** Urban forest soil infiltration, affected by various factors, is closely related with surface runoff. This paper studied the effect of urban forest types, vegetation configuration and soil properties on soil infiltration. In our study, 191 typical plots were sampled in Changchun City, China to investigate the soil infiltration characteristics of urban forest and its influencing factors. Our results showed that the steady infiltration rates of urban forest soil were highly variable. High variations in the final infiltration rates were observed for different vegetation patterns and compaction degrees. Trees with shrubs and grasses had the highest infiltration rate and trees with bare land had the lowest infiltration rate. In addition, our results showed that the soil infiltration rate decreased with an increase in the bulk density and with a reduction in the soil organic matter content and non-capillary porosity. The soil infiltration rate also had significantly positive relationships with the total porosity and saturated soil water content. Urban soil compaction contributed to low soil infiltration rates. To increase the infiltration rate and water storage volume of urban forest soil, proper techniques to minimize and mitigate soil compaction should be used. These findings can provide useful information for urban planners about how to maximize the water volume of urban forest soil and decrease urban instantaneous flooding.

**Keywords:** soil infiltration ; urbanization gradients; urban forest; plant communities; soil property

**Citation:** WANG Peijiang, ZHENG Haifeng, REN Zhibin, ZHANG Dan, ZHAI Chang, MAO Zhixia, TANG Ze, HE Xingyuan, 2018. Effects of Urbanization, Soil Property and Vegetation Configuration on Soil Infiltration of Urban Forest in Changchun, Northeast China. *Chinese Geographical Science*, 28(3): 482–494. <https://doi.org/10.1007/s11769-018-0953-7>

## 1 Introduction

With rapid urbanization, natural vegetated soils are replaced with impervious surfaces. This land conversion could exert profound influences on hydrological processes such as inhibiting rainwater infiltration and increasing surface runoff and peak discharge rates (Hood et al., 2007; Asadian and Weiler, 2009; Endreny and Collins, 2009; Jia et al., 2015; Zhang Biao

et al., 2015a). Excessive runoff accompanied by low urban forest coverage could frequently cause flooding in the urban areas and pose threats to life and property (Pataki et al., 2011; Yao et al., 2015). The increase in flood risk due to short-term heavy rains has been a major concern in many cities (United Nations, 2012). Nearly all cities have set up mitigation strategies to improve the design standards of urban drainage pipe networks. However, such strategies are costly and in-

Received date: 2017-01-09; accepted date: 2017-05-02

Foundation item: Under the auspices of Excellent Young Scholars of Northeast Institute of Geography and Agroecology, Chinese Academy of Sciences (No. DLSYQ 13004), Chinese Academy of Sciences/State Administration of Foreign Experts Affairs International Partnership Program for Creative Research Teams (No. KZZD-EW-TZ-07-09), Strategic Priority Research Program of Chinese Academy of Sciences (No. KFZD-SW-302-03)

Corresponding author: HE Xingyuan. E-mail: [hexingyuan196209@163.com](mailto:hexingyuan196209@163.com)

© Science Press, Northeast Institute of Geography and Agroecology, CAS and Springer-Verlag GmbH Germany, part of Springer Nature 2018

crease the pressure on sewage treatment plants. Therefore, effective and low-cost measures to solve water-logging in cities need to be explored.

The effects of vegetation on surface water runoff have been extensively studied (Sanders, 1986; Joffre and Rambal, 1993; Ellis *et al.*, 2006). Much research has focused on natural vegetation of the watershed in rural areas (Bu *et al.*, 2010; Armson *et al.*, 2013; Wang *et al.*, 2013). There have been fewer studies about runoff mitigation of urban vegetation, especially for urban forests. Urban forests, mainly consisting of trees, shrubs and lawns, positively influence urban hydrology through soil water storage and enhanced infiltration into the root and soil zones, as well as interception by the canopy and plant stems (Gill *et al.*, 2007; Bartens *et al.*, 2008; Zhang *et al.*, 2012). Urban forest soil infiltration was closely related with rainfall and irrigation water penetrating soils and then migrating and being stored in the soil (Kang *et al.*, 2008; Yang and Zhang, 2008; Cheng *et al.*, 2009; Saito *et al.*, 2016). It has significant influence on surface runoff and water management (Alaoui and Goetz, 2008; Cheng *et al.*, 2017; Zhang *et al.*, 2017). Although many studies on urban forest soils have been conducted, most of them have focused only on soil fertility (Shan *et al.*, 2009; Li *et al.*, 2015), heavy metal pollutants (Pan *et al.*, 2008; Fang *et al.*, 2012) and soil microorganisms (Zhao and Guo, 2010; Chen *et al.*, 2012). Most previous studies in urban areas primarily used the results from rural areas to generate the runoff reduction values (Yao *et al.*, 2015; Zhang Biao *et al.*, 2015). This method did not take into account the impact that the urban conditions may have on urban infiltration. In addition, urbanization has dramatically alerted the urban soil through sealing, intensive disturbance, deposition of building and daily rubbish, sedimentation of air dust, and infiltration of sewage (Zhao *et al.*, 2007; Shaw *et al.*, 2010). These processes damage and change the urban soil compared with the natural soil. It is important to understand scientifically the effects of urbanization on soil infiltration. Urbanization effects are typically evaluated along urban-rural gradients, which have been used in studies considering soil properties, forest soil heavy metals, polycyclic aromatic hydrocarbons and soil organic carbon (Pouyat and McDonnell, 1991; Pouyat *et al.*, 2002; Wong *et al.*, 2004; Zhang, 2004; Pouyat *et al.*, 2008). Many results showed that soil organic matter, water-stable aggregates, cation exchange

capacity, total nitrogen, total phosphorus and heavy metal concentrations (such as lead, copper, and nickel) increased from rural to urban zones (Zhang *et al.*, 2003; Lu *et al.*, 2009). In China, cities are usually sprawled through ring road development and different ring roads often represent the urbanization gradients (Huang *et al.*, 2010; Zhai *et al.*, 2017). Therefore, studying the urbanization effects on forest soil infiltration through the use of ring road-related gradients is available. Vegetation can improve the soil properties in various ways depending on different community composition (Bezemer *et al.*, 2006; Yimer *et al.*, 2006; Fu *et al.*, 2010; Oldfield *et al.*, 2014; Regüés *et al.*, 2017). Many results showed that different vegetation composition had different impact on soil organic carbon storage, soil microbial community and soil quality (Guo and Gifford, 2002; Hooker and Compton, 2003; Gao *et al.*, 2009; Merilä *et al.*, 2010; Hou *et al.*, 2012). However, there have been fewer studies on the effect of different vegetation composition on soil properties, especially in cities.

Until now, only a few studies on urban green space soil infiltration rates and influencing factors have been conducted (Winzig, 2000; Yang and Zhang, 2011; Wei *et al.*, 2012), with less emphasis on soil infiltration characteristics of different forest types, different plant communities and different urbanization gradients. Changchun City in Northeast China has undergone rapid urbanization over the past decades. The urban vegetation coverage and soil have changed notably. Understanding the spatial distribution of soil infiltration of urban forests in Changchun and its relationships with urbanization are critical for planning a no-flood and sustainable city. Understanding the current situation of soil infiltration and its influencing factors can provide the theory base for urban forest management. The aims of this work were as follows: 1) describe the urban forest soil infiltration characteristics, 2) study the effect of urban forest types, urbanization gradients and plant communities on soil infiltration, and 3) discuss its influencing factors.

## 2 Materials and Methods

### 2.1 Study area

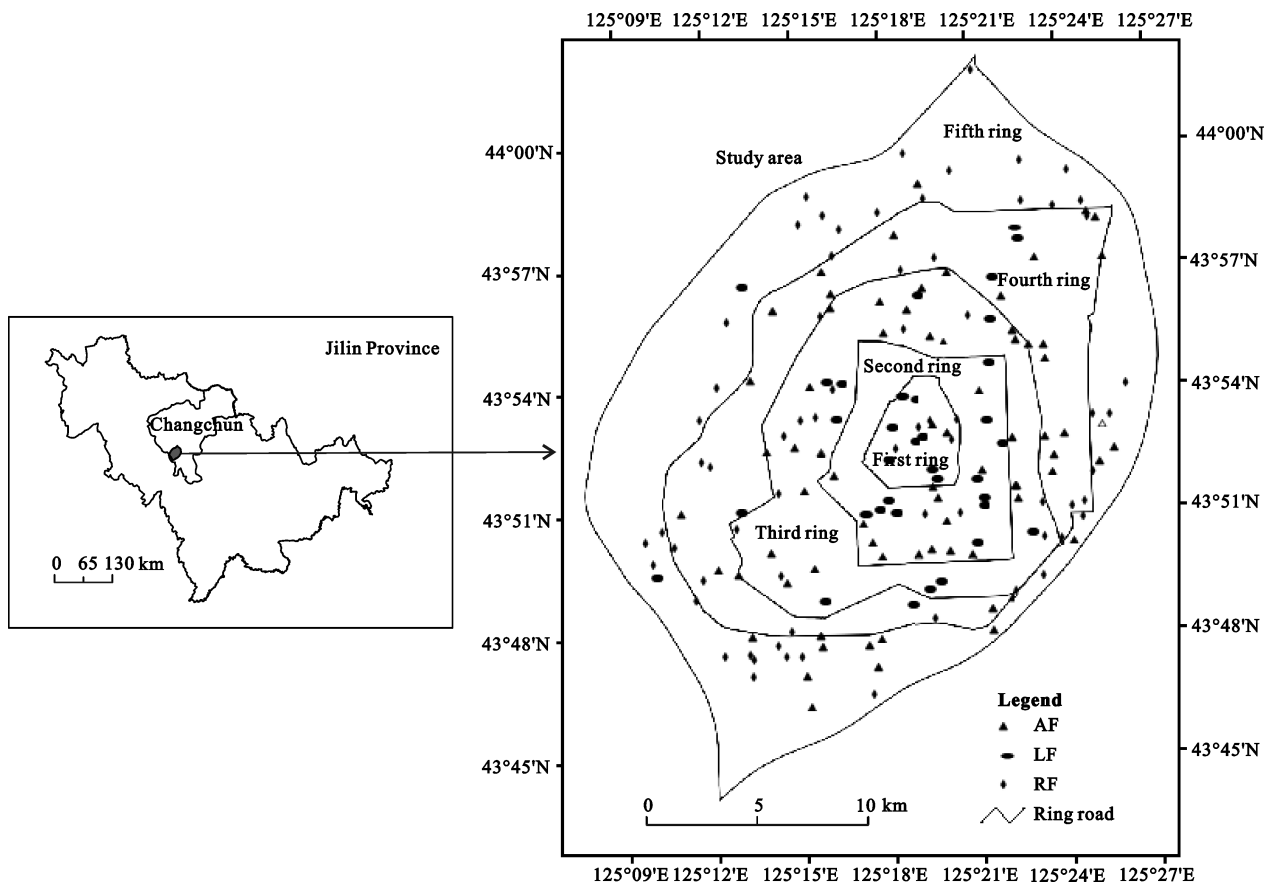
Changchun (43°05′–45°15′N, 124°18′–127°02′E), the capital of Jilin Province, is located in the northeastern plain of China. This region is characterized by a subhu-

mid climate with a continental monsoon (Ren et al., 2013; Zhang Dan et al., 2015). The mean annual temperature is 4.8°C and the mean annual precipitation is 569.6 mm. Nearly 70% of the annual precipitation is concentrated in the summer season. The city has witnessed a rapid urbanization and economic growth over the last twenty years. Its urban population has accelerated notably in recent decades with urban expansion that followed the form of concentric ring roads. The main urban area of Changchun had increased from 117 to 365 km<sup>2</sup>, and the population increased from 2.20 million to 3.62 million during the period from 1994 to 2010 (Huang et al., 2009; Li et al., 2012). Urban waterlogging caused by the heavy precipitation and old urban drainage piping systems has become more frequent in recent years (Xiao et al., 1989; Xi et al., 2016). The geology of the region consists of Quaternary alluvial-diluvial deposits and the natural soil types in the study area are, mainly black soil, dark brown soil and meadow soil according to the Classification and codes for Chinese soil (National standard, GB/T 17296-2009) (Yang et al., 2011). The urban forest cover of the main urban area is

106.81 km<sup>2</sup>, mainly composed of coniferous forest and broad-leaved forest (Zhang Dan et al., 2015). The common trees in Changchun includes *Armeniaca mandshurica*, *Pinus sylvestris* var. *mongolica*, *Pinus tabulaeformis*, *Larix gmelinii*, *Salix babulonica*, *Populus alba* and *Ulmus pumila* (Zhai et al., 2017). The understory shrubs and grasses includes *Swida alba*, *Spiraea salicifolia*, *Amygdalus triloba* and *Poa annua*.

## 2.2 Soil sampling

The sampling area (43°43'–44°04'N, 125°07'–125°27'E) is primarily within the fifth ring road of Changchun, with an area of 524.06 km<sup>2</sup>. Based on location, function and management objectives, urban forests are classified into five types (He, 2004): road forest (RF), attached forest (AF), landscape and relaxation forest (LF), ecological and public welfare forest (EF) and production and management forest (PF). A total of 191 representative sites were surveyed in Changchun (Fig. 1), which included 72 plots for RF, 73 plots for AF, and 46 plots for LF. According to the understory vegetation, the plots were divided into four types: trees with shrubs and



**Fig. 1** Location of study area and sampling sites. RF, road forest; AF, attached forest; LF, landscape and relaxation forest.

grasses (FSG, 64 plots), trees with grasses (FG, 86 plots), trees with sparse grasses (FFG, 17 plots) and trees with bare land (FB, 24 plots). Field surveys were conducted from June to August in 2014. Each plant community plot was 400 m<sup>2</sup>. Both the widths and lengths of the plots were normally 20 m and adjusted according to the situation. Urban soil compaction occurs primarily in the soil surface (Dornauf and Burghardt, 2000; Yang et al., 2005); thus, soils were sampled at a depth of 0–10 cm. Undisturbed soil cores were collected with standard core rings (height, 52.00 mm; diameter, 70.00 mm) to determine the soil water content, bulk density, and porosity. In addition, samples were taken from a depth of 0–10 cm to analyze the pH, organic matter and other properties. Each sample was replicated three times.

### 2.3 Soil property analysis

Soil physical and chemical properties were determined according to standard methods (Institute of Soil Science, Chinese Academy of Sciences, 1978; Lu, 2000). The antecedent soil water content (AWC) was determined by the oven drying method, and the bulk density (BD) was obtained from soil core rings. The total porosity (TP) was calculated from the bulk density and the particle density ( $d_s=2.65$  g/cm<sup>3</sup>) (Equation (1)), the capillary porosity (CP) was obtained from the intact soil core (Equation (2)), and the non-capillary porosity (NCP) was calculated from the total porosity and the capillary porosity (Equation (3)). Saturated soil water content (SSWC) (g H<sub>2</sub>O/100 g dry soil) was determined by mass loss from saturated soil samples during oven drying at 105°C to constant weight, and then it was calculated on a volumetric basis using the determined BD value. The soil pH was obtained with a glass electrode at a 1 : 2.5 sample/water ratio. Soil organic matter content (SOC) was determined by acid-dichromate digestion. Aggregate stability (WAS) was determined using an adaptation of the wet sieving method (Yoder, 1936).

$$TP = \left(1 - \frac{BD}{d_s}\right) \times 100 \quad (1)$$

$$CP = W_c \times \frac{BD}{V} \times 100 \quad (2)$$

$$NCP = TP - CP \quad (3)$$

where  $TP$  is the total soil porosity (%),  $BD$  is the soil

bulk density (g/cm<sup>3</sup>), and  $d_s$  is the soil particle density (g/cm<sup>3</sup>),  $CP$  is the soil capillary porosity (%),  $NCP$  is the soil non-capillary porosity (%),  $W_c$  is the soil capillary water content (%), and  $V$  is the volume of the soil core (cm<sup>3</sup>).

The initial and final infiltration rate (cm/h) of soil was determined by the dual-ring method (Liu et al., 2007). According to the process of soil infiltration, we chose the mean infiltration rate for the first 3 min as the initial infiltration rate (IIR), and then chose the average infiltration rate for 0–15 min as the average infiltration rate of the stage I (AIRSI). Likewise, the average infiltration rate of stage II (AIRSII) was for the period of 15–30 min and the average infiltration rate of stage III (AIRSIII) was for the period from 30 to 60 min. The average infiltration rate of the final 3 min was taken as the steady infiltration rate (SIR). The average infiltration rate of the 0–60 min period was chosen as the overall average infiltration rate (AIR) (Wu et al., 2016). According to the criteria proposed by Kohnke (1968), the steady infiltration rates were categorized into seven levels: very slow ( $k_s \leq 0.1$  cm/h), slow ( $0.1 < k_s \leq 0.5$  cm/h), slow to medium ( $0.5 < k_s \leq 2.0$  cm/h), medium ( $2.0 < k_s \leq 6.3$  cm/h), medium to fast ( $6.3 < k_s \leq 12.7$  cm/h), fast ( $12.7 < k_s \leq 25.4$  cm/h) and very fast ( $k_s > 25.4$  cm/h).

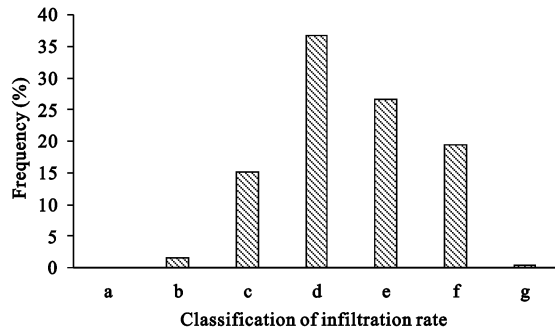
### 2.4 Data Analysis

The values of soil property were averaged for three samples from each plot. Pearson correlation coefficients were calculated to detect relationships between soil properties and infiltration variables. One-way analysis of variance (ANOVA;  $P < 0.05$ ) was used to compare the differences between means. Principal component analysis (PCA) was used to determine the influencing factors. The SPSS 17.0 was used to carry out the analysis.

## 3 Results

### 3.1 Steady infiltration rate level

In our study, the steady infiltration rates were categorized into seven levels (Fig. 2). The steady infiltration rates of urban forest soils varied widely and ranged from 0.23 to 25.50 cm/h. The majority of infiltration rates were categorized as slow to medium, medium, medium to fast and fast. No very slow infiltration rate soil was in Changchun urban forest. Slow and very fast infiltration rates accounted for 2.00% of the samples. Medium



**Fig. 2** Frequency of distribution of steady infiltration rate. a=very slow; b=slow; c=slow to medium; d=medium; e=medium to fast; f=fast; g=very fast

infiltration rate accounted for 37.00% of the samples and nearly 53.40% of the soil infiltration rates were lower than 6.30 cm/h (medium to fast infiltration rate).

### 3.2 Soil infiltration rate of different vegetation types beneath forests and different urban forest

As shown in Table 1, infiltration rate began with high values and decreased gradually to a final constant rate. The infiltration rates varied widely over infiltration stages. Moreover, the difference between the initial and

final infiltration rate was large. The RF had higher mean values of IIR, SIR, AIR, AIRSI, AIRSII and AIRSIII than other urban forest types, and LF had the lowest values. The FSG had the highest values and FB had the lowest values. In our study, the infiltration rate of FFG was higher than FG in AF and LF but not in RF. As shown in Fig. 3A, the final infiltration rates of RF, LF and AF were not significantly different because large variations were observed in these urban forest types. However, significant differences in the infiltration rates were observed between FSG and FB (Fig. 3B). Compared with FB, the infiltration rates of FSG improved by 45.95%–102.18%. As shown in Table 1, trees with grasses or shrubs had higher values of AIR, AIRSI, AIRSII, AIRSIII and IIR than the trees with bare land under them.

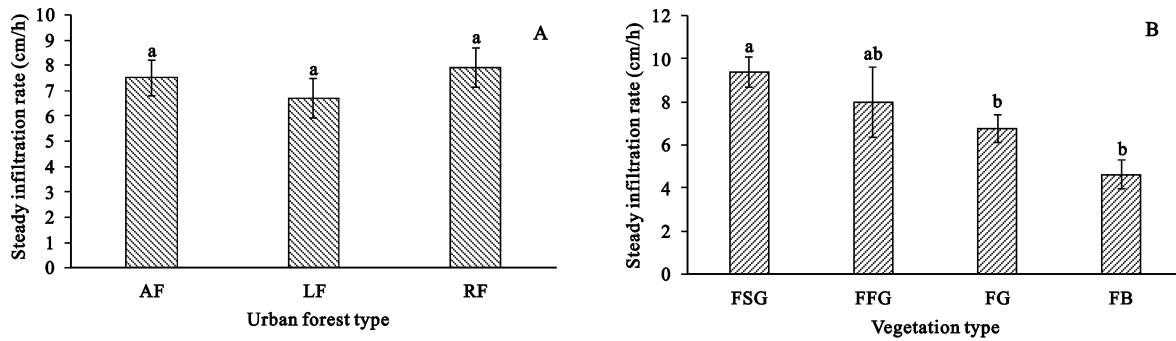
### 3.3 Soil infiltration rate of gradients of urbanization

As shown in Fig. 4, the soil infiltration rate of LF in the second ring was the lowest and that in the third ring was the highest. The soil infiltration rate of RF in the first ring was the lowest and in the fourth ring was the highest. In addition, there were no significant differences

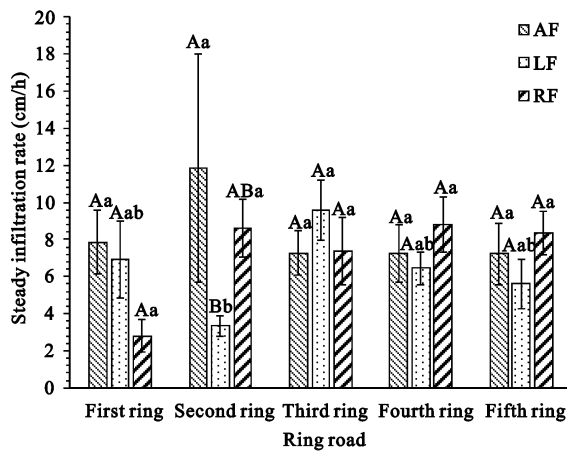
**Table 1** Soil infiltration rates of urban forest soil

Type	Vegetation	IIR (cm/h)	SIR (cm/h)	AIR (cm/h)	AIRSI (cm/h)	AIRSII (cm/h)	AIRSIII (cm/h)
AF	FSG	24.34±20.81	9.57±6.14	12.01±9.07	17.76±14.68	12.85±9.49	10.02±6.45
	FG	15.21±20.09	5.84±5.65	7.47±7.81	10.23±12.67	7.12±7.40	5.89±5.95
	FFG	20.83±10.71	11.20±6.32	13.15±7.48	16.57±9.43	13.08±7.69	11.48±6.47
	FB	12.15±8.56	4.67±2.39	6.16±3.11	9.07±6.02	5.72±2.84	4.92±2.41
	MV	18.98±19.86	7.51±5.99	9.48±8.36	13.55±13.46	9.63±8.55	7.75±6.32
LF	FSG	22.88±19.79	9.22±4.97	10.74±8.12	15.83±11.44	11.23±6.69	9.61±5.49
	FG	14.15±10.80	6.56±5.22	7.98±6.02	10.80±8.58	8.21±6.55	6.93±5.44
	FFG	15.94±15.30	7.58±8.58	8.97±9.75	11.50±11.83	8.67±9.50	7.85±9.02
	FB	11.22±10.39	4.16±2.64	5.19±3.48	7.00±5.48	5.01±3.13	4.37±2.67
	MV	15.31±13.12	6.70±5.23	8.06±6.43	11.08±9.07	8.22±6.47	7.03±5.49
RF	FSG	22.84±21.72	9.17±6.68	11.35±8.41	14.99±11.85	11.17±8.30	9.51±6.96
	FG	20.03±22.72	8.20±7.39	10.66±9.91	14.65±15.61	10.34±9.62	8.72±7.83
	FFG	15.74±11.98	7.19±6.47	8.99±7.95	11.65±10.33	8.99±8.15	7.67±6.83
	FB	13.51±15.61	4.96±3.99	6.62±6.56	9.30±10.98	6.50±6.39	5.34±4.57
	MV	19.41±20.13	7.90±6.62	10.05±8.68	13.53±12.97	9.85±8.52	8.33±6.99

Notes: AF, attached forest; LF, landscape and relaxation forest; RF, road forest. MV is the average infiltration rates of different forest types. FSG, trees with shrubs and grasses; FG, trees with grasses; FFG, trees with sparse grasses; FB, trees with bare land. IIR is the mean infiltration rate for the first 3 min; AIRSI is the average infiltration rate of stage I for the period of 0–15 min; AIRSII is the average infiltration rate of stage II for the period of 15–30 min; AIRSIII is the average infiltration rate of stage III for the period of 30–60 min; SIR is the average infiltration rate of the final 3 min and AIR is the average infiltration rate of the 0–60 min period.



**Fig. 3** Final infiltration rate for different urban forests (A) and vegetation types (B). Error bars indicate the standard error, a and b represent significant differences at a confidence level of 95%. AF, attached forest; LF, landscape and relaxation forest; RF, road forest. FSG, trees with shrubs and grasses; FFG, trees with sparse grasses; FG, trees with grasses; FB, trees with bare land.



**Fig. 4** Mean infiltration rates of different forest types in different urbanization gradients. A and B represent significant differences of different forest types in the same ring at the confidence level of 95%; a and b represent significant differences of the same forest types in the different rings at a confidence level of 95%. AF, attached forest; LF, landscape and relaxation forest; RF, road forest.

among the mean soil infiltration rates of AF and RF in

the five rings. However, significant differences were observed among the mean soil infiltration rates of LF in the five rings. There were significant differences among the mean soil infiltration rates of these three types of urban forest just in the second ring. In the second ring, LF had the highest soil infiltration rate and soil infiltration rate of AF was close to RF.

### 3.4 Influencing factors of soil infiltration

In our study, the infiltration rates were correlated to the bulk density, total porosity, capillary porosity and saturated soil water content (Table 2). There were significant negative relationships between SIR and BD and significant positive relationships between SIR and TP, CP and SSWC. Principal component analysis (PCA) was applied to the aforementioned factors. As shown in Table 3, the eigenvalues of the first four PCs were great than 1 and the four components contributed to more than 77.79% of the total variance. The highly weighted parameters in PC1 were BD, TP, CP, AWC and SSWC. The BD had the highest loading value and was strongly

**Table 2** Correlation matrix between infiltration rates and soil properties

Soil infiltration rate	pH	EC	BD	AWC	TP	CP	NCP	WAS	SOC	SSWC
IIR	-0.007	0.063	<b>-0.191**</b>	0.068	0.191**	0.166**	0.021	0.088	0.059	0.141
SIR	0.031	0.067	<b>-0.300**</b>	0.124	0.300**	0.245**	0.061	0.074	0.033	0.188**
AIR	0.013	0.071	<b>-0.296**</b>	0.122	0.296**	0.240**	0.063	0.108	0.060	0.214**
ARISI	-0.007	0.063	<b>-0.260**</b>	0.103	0.260**	0.224**	0.033	0.092	0.058	0.186*
ARISII	0.011	0.070	<b>-0.310**</b>	0.132	0.310**	0.261**	0.049	0.094	0.053	0.205**
ARISIII	0.033	0.079	<b>-0.313**</b>	0.131	0.313**	0.257**	0.063	0.076	0.047	0.195**

Notes: \* indicates significant correlation and \*\* indicates highly significant correlation. EC, electrical conductivity; BD, bulk density; AWC, antecedent soil water content; TP, total porosity; CP, capillary porosity; NCP, non-capillary porosity; WAS, aggregate stability; SOC, soil organic matter content; SSWC, saturated soil water content; IIR is the mean infiltration rate for the first 3 min; AIRSI is the average infiltration rate of stage I for the period of 0–15 min; AIRSII is the average infiltration rate of stage II for the period of 15–30 min; AIRSIII is the average infiltration rate of stage III for the period of 30–60 min; SIR is the average infiltration rate of the final 3 min and AIR is the average infiltration rate of the 0–60 min period.

**Table 3** Extracted components and contributing factors of infiltration rates by principal component analysis

Eigenvector	Principal component			
	PC1	PC2	PC3	PC4
BD	-0.947	-0.093	0.152	-0.061
TP	0.947	0.093	-0.152	0.061
CP	0.887	-0.252	0.173	-0.236
NCP	0.030	0.617	-0.574	0.531
pH	0.188	-0.487	0.160	0.456
EC	0.203	0.596	0.451	-0.116
SSWC	0.845	-0.144	0.016	0.031
WAS	-0.039	-0.018	0.591	0.657
AWC	0.775	-0.125	-0.083	0.066
SOC	0.306	0.708	0.036	-0.117
Eigenvalue	4.072	1.557	1.122	1.011
Cumulative variance explained (%)	40.660	56.580	67.610	77.790

Notes: PC, principal component; BD, bulk density; TP, total porosity; CP, capillary porosity; NCP, non-capillary porosity; EC, electrical conductivity; SSWC, saturated soil water content; WAS, aggregate stability; AWC, antecedent soil water content; SOC, soil organic matter content.

correlated with TP, CP, AWC and SSWC (Table 4). Thus, TP, CP, AWC and SSWC were eliminated, while BD was retained as a parameter that affected the soil infiltration capacity. In PC2, the highly loaded parameters included NCP, EC and SOC. The SOC had the highest loading value and was strongly correlated with NCP ( $r=0.154$ , Table 4) and EC ( $r=0.405$ , Table 4). Therefore, SOC was selected. In PC3 and PC4, NCP and WAS had the highest loading values. No significant correlation was found between NCP and WAS. So, NCP

and WAS were chosen. In summary, the most critical parameters, as determined from the PCA, that would affect soil infiltration capacity were BD, NCP, WAS and SOC.

## 4 Discussion

### 4.1 Impact of vegetation cover on soil infiltration

A steady infiltration rate, equivalent to saturated hydraulic conductivity, is one of the important indicators to evaluate soil physical characteristics and a critical component of most urban runoff models. Soil infiltration is affected by various factors such as vegetation cover, soil texture and structure, soil organic matter, antecedent water content, management system, disturbance age and landscape position (Radke and Berry, 1993; Pitt et al., 1999; Kumar et al., 2012). Many previous studies have showed that there was a great difference in the soil infiltration rate of different land use and different functional urban green areas (Yang et al., 2008; Wei et al., 2012; Chen et al., 2014). There was wide variability in soil infiltration rates among different urban forest types in our study. Vegetation characteristics have important effects on soil infiltration capacity. In our study, the FSG had the highest values followed by FG, and FB had the lowest values (Fig. 3B). The infiltration rates in areas with shrubs and grasses were relatively higher because shrubs and grasses can loosen compacted soil and can aid in the formation of soil macropores and good soil structure (Meek et al., 1989; Wu et al., 2016). In addition, shrubs protected the soil

**Table 4** Correlation matrix among the soil properties

Soil property	BD	TP	CP	NCP	pH	EC	AWC	SSWC	WAS	SOC
BD	1	-1.000**	-0.844**	-0.177*	-0.116	-0.172*	-0.635**	-0.688**	0.048	-0.226**
TP		1	0.844**	0.177**	0.116	0.172*	0.635**	0.688**	-0.048	0.226**
CP			1	-0.379**	0.176*	0.123	0.592**	0.700**	-0.031	0.128
NCP				1	-0.122	0.071	0.008	-0.096	-0.026	0.154*
pH					1	0.011	0.163*	0.213**	0.071	-0.191**
EC						1	0.076	0.052	-0.001	0.405**
AWC							1	0.700**	-0.030	0.116
SSWC								1	0.008	0.226**
WAS									1	0.059
SOC										1

Notes: BD, bulk density; TP, total porosity; CP, capillary porosity; NCP, non-capillary porosity; EC, electrical conductivity; SSWC, saturated soil water content; WAS, aggregate stability; AWC, antecedent soil water content; SOC, soil organic matter content.

from compaction. Alternatively, the infiltration rate was very low in urban forests with bare land because the soil was apt to be trampled by pedestrians. The infiltration rate of FFG was higher than FG in AF and LF, possibly because the infiltration capacity is decreased by grasses (Archer *et al.*, 2002; Fischer *et al.*, 2014). Thus, shrubs are better than grasses in the improvement of soil structure in urban greening. Therefore, in order to make full use of urban green space soil for rainwater reduction, a proper vegetation configuration is needed. For example, in urban greening, more shrubs and less lawns should be promoted.

#### 4.2 Impact of urbanization on soil infiltration

In the past few decades, China has experienced a rapid and unprecedented process of urbanization, with urban areas expanding almost exponentially outwards in many cities in parallel with ring road construction. In China, different ring roads often represent the urbanization gradients (Huang *et al.*, 2010; Zhai *et al.*, 2017). During the process of urbanization, forests and soils were significantly influenced by human activities and large areas of urban forests have been occupied due to accelerating urbanization. Meanwhile, many afforestation movements such as ‘forest city’ or ‘eco-city’ have been initiated in cities due to citizens’ desire for a better quality of life (Bae and Ryu, 2015; Lv *et al.*, 2016). Urban forest soil is affected not only by plant-soil interaction and afforestation but also by human activities such as trampling (Jim, 1998c; Hobbie *et al.*, 2007). Our results indicated that there were no obvious differences in the steady infiltration rate of AF and RF among different ring roads. The steady infiltration rate of LF in the third ring was significantly higher than other rings. One reason for such differences might be that LF is mainly located in the semi-natural environments (such as public parks, forest parks, and historic sites). LF in the third ring were less disturbed, LF in the first ring and the second ring were intensively disturbed by people but LF in the fourth and fifth are under construction. Compared with LF, AF and RF are more closely impacted by humans. Therefore, the interference degree on AF and RF soil of different urban gradients might not differ greatly. To sum up, the impact of urbanization on urban soil infiltration is complex, and the mechanisms that drive this difference due to the complex factors controlling soil infiltration in urbanized regions still needs to be ex-

plored.

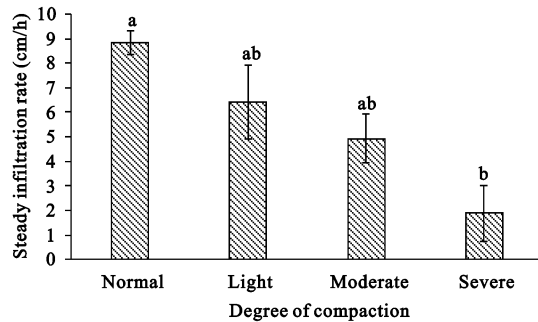
#### 4.3 Impact of soil properties on soil infiltration

Soil water infiltration is a key process in the water cycle since it controls, *inter alia*, the surface water-groundwater relationship. The soil properties play a crucial role in this process. Many studies have demonstrated that the infiltration rate depends mainly on soil properties, such as initial moisture content, hydraulic conductivity, soil texture, porosity, swelling degree of soil colloids, organic matter content, and chemical properties (Saxton *et al.*, 1986; Bagarello and Sgroi, 2004; Chai *et al.*, 2007; Bormann and Klaassen, 2008; Fischer *et al.*, 2014). Soil bulk density is one of the most important factors that influence infiltration capacity (Winzig, 2000; Yang *et al.*, 2008). Our results also showed that the soil infiltration rate decreased with increasing bulk density. For example, at a soil bulk density of  $1.44 \text{ g/cm}^3$ , the infiltration rate was lower than  $5.00 \text{ cm/h}$ . In contrast, the bulk densities of soils that had higher infiltration rates were  $1.19\text{--}1.35 \text{ g/cm}^3$ .

Soil pores provide channels for water movement in the soil. If there is more connectivity between pores, the soil infiltration rate will be increased. Past studies have found that soil infiltration rates had significant positive linear correlations with the total porosity and non-capillary porosity (Yang and Zhang, 2011; Li *et al.*, 2013). Our results showed that there were positive relationships between infiltration rates and soil total porosity and non-capillary porosity. Urban soils are frequently mixed with many gravel, coal cinders, construction waste and other artificial substances in city greening. The existence of gravel creates a number of macropores in soil, which easily form preferential flow that is mainly influenced by gravity and not capillary pores (Yang *et al.*, 2008).

Compaction is the most serious form of the physical degradation in urban areas. Compaction usually leads to increases in bulk density and decreases in porosity and formation of a crust layer that prevents water infiltration into the surface soil (Bruand and Cousin, 1995; Jim, 1998a; Richard *et al.*, 2001). In our study, the final infiltration rate decreased with increased soil compaction (Fig. 5). Furthermore, the final infiltration rate of non-compacted soil was significantly different from that of severely compacted soils. Specifically, the infiltration rate of non-compacted soil was  $8.84 \text{ cm/h}$ , and the infil-





**Fig. 5** Final infiltration rates of urban forest soil with different degrees of compaction. Error bars indicate the standard error, a and b represent significant differences at a confidence level of 95.00%. According to Yang and Zhang (2011), urban forest soil of Changchun can be categorized into five compaction levels: normal,  $d_s < 1.30 \text{ g/cm}^3$ ; light,  $1.30 \text{ g/cm}^3 \leq d_s < 1.40 \text{ g/cm}^3$ ; moderate,  $1.40 \text{ g/cm}^3 \leq d_s < 1.50 \text{ g/cm}^3$ ; severe,  $1.50 \text{ g/cm}^3 \leq d_s < 1.60 \text{ g/cm}^3$ ; extreme,  $d_s \geq 1.60 \text{ g/cm}^3$

tration rate of the severely compacted soil was 1.88 cm/h. Generally, urban green space soils are compacted mainly by mechanical compaction in urban greening and human trampling. Compaction affects root growth of the plant and inhibits water penetration into the soil, causing waterlogging in urban areas (Yang and Zhang, 2011). Therefore, in the process of urban greening the measures should be taken to loosen the soil and avoid compaction.

In the process of urbanization, soils are typically degraded by a wide range of modifications including vegetation clearing, topsoil removal, grading, and compaction. These practices significantly influence soil physical characteristics (Jim, 1993, 1998b; Alaoui et al., 2011) and ultimately lead to the loss of critical soil-mediated ecosystem services such as storm water mitigation (Pitt et al., 2008), carbon storage (Chen et al., 2013), and net primary productivity (Milesi et al., 2003). Many studies showed that soil organic amendments could help the formation of soil aggregates and improve the soil structure and therefore has a significant effect on the increasing infiltration rates (Boyle et al., 1989; Martens and Frankenberger, 1992; Arvidsson, 1998; Celik et al., 2010; Brown and Cotton, 2011). In our study, there were positive relationships between soil organic matter and infiltration rate. Thus, in the process of urban green spaces management and maintenance, returning the litter to the green land, increasing the surface organic coverage or organic manure fertilizer to increase the infil-

tration rate should be advocated.

#### 4.4 Comparison of urban greenspace soil infiltration rate with other cities

Compared with other studies in China, the percentage of medium soil infiltration rate in Changchun (37%) was similar to Hefei (Wei et al., 2012) and Shanghai (Nie et al., 2008) greenspaces. The percentage of soil infiltration rate less than medium to fast in Changchun (53.40%) was similar to Nanjing (Yang et al., 2008) and lower than Hefei city (78.90%) and Shanghai city (84.21%) (Wei et al., 2011). The soil infiltration rates of Changchun urban forest ranged from 0.23 to 25.50 cm/h. Soil infiltration rates of residential lawns in central and south-central Pennsylvania ranged from 0.40 to 10.00 cm/h and 0 to 41.80 cm/h, respectively (Hamilton and Waddington, 1999; Woltemade, 2010). The urban soil infiltration rates in Oconomowoc, Wisconsin, ranged from 0 to 38 cm/h (Pitt et al., 1999) and in Montgomery County, Virginia, ranged from 0.40 to 2.30 cm/h (Chen et al., 2014). Thus, urban soil infiltration rates varied largely for soils from different regions. Soil texture differed due to development under different soil parent materials (Bormann and Klaassen, 2008). Even in the same city there may be some inhomogeneous soils. Moreover, green space management, area and distribution also differed from one city to another. These all might have impact on soil infiltration.

## 5 Conclusions

In this study, detailed field experiments were conducted using double ring infiltrometers to examine the effects of urban forest type and community composition on soil infiltration rates within the fifth ring road of Changchun. The steady infiltration rates of urban forest soil were highly variable and ranged from 0.23 to 25.50 cm/h. Almost 50% of the soil infiltration rates were very low and it showed that the urban forest soil infiltration capacity of Changchun was worse. Soil infiltration rates of road forest in the infiltration process were higher than attached forest and landscape and relaxation forest. Urbanization had different influences on soil infiltration of different urban forest types. Landscape and relaxation forest soil which was strongly disturbed by the local people had the lower infiltration rate. The impact of urbanization on soil infiltration of the attached forest and

road forest might be not obvious. Large variations in the final infiltration rate were observed with different vegetation patterns. Trees with shrubs and grasses had the highest infiltration rates and trees. Large variations in the final infiltration rates of urban forest soils were attributed to the effects of soil bulk density and porosity due to high human activity. Thus, the degree of compaction was the most important factor that affected the urban forest soil infiltration rate. Therefore, in urban green spaces construction and management, mechanical compaction and human trampling must be avoided, and proper vegetation configuration will improve the role of urban forest soil in mitigation runoff.

## References

- Alaoui A, Goetz B, 2008. Dye tracer and infiltration experiments to investigate macropore flow. *Geoderma*, 144(1–2): 279–286. doi: 10.1016/j.geoderma.2007.11.020
- Alaoui A, Lipiec J, Gerke H H, 2011. A review of the changes in the soil pore system due to soil deformation: a hydrodynamic perspective. *Soil and Tillage Research*, 115–116: 1–15. doi: 10.1016/j.still.2011.06.002
- Archer N A L, Quinton J N, Hess T M, 2002. Below-ground relationships of soil texture, roots and hydraulic conductivity in two-phase mosaic vegetation in South-east Spain. *Journal of Arid Environments*, 52(4): 535–553. doi: 10.1006/jare.2002.1011
- Armson D, Stringer P, Ennos A R, 2013. The effect of street trees and amenity grass on urban surface water runoff in Manchester, UK. *Urban Forestry & Urban Greening*, 12(3): 282–286. doi: 10.1016/j.ufug.2013.04.001
- Arvidsson J, 1998. Influence of soil texture and organic matter content on bulk density, air content, compression index and crop yield in field and laboratory compression experiments. *Soil and Tillage Research*, 49(1–2): 159–170. doi: 10.1016/S0167-1987(98)00164-0
- Asadian Y, Weiler M, 2009. A new approach in measuring rainfall interception by urban trees in coastal British Columbia. *Water Quality Research Journal of Canada*, 44(1): 16–25.
- Bae J, Ryu Y, 2015. Land use and land cover changes explain spatial and temporal variations of the soil organic carbon stocks in a constructed urban park. *Landscape and Urban Planning*, 136: 57–67. doi: 10.1016/j.landurbplan.2014.11.015
- Bagarello V, Sgroi A, 2004. Using the single-ring infiltrometer method to detect temporal changes in surface soil field-saturated hydraulic conductivity. *Soil and Tillage Research*, 76(1): 13–24. doi: 10.1016/j.still.2003.08.008
- Bartens J, Day S D, Harris J R et al., 2008. Can urban tree roots improve infiltration through compacted subsoils for stormwater management? *Journal of Environmental Quality*, 37(6): 2048–2057. doi: 10.2134/jeq.2008.0117
- Bezemer T M, Lawson C S, Hedlund K et al., 2006. Plant species and functional group effects on abiotic and microbial soil properties and plant-soil feedback responses in two grasslands. *Journal of Ecology*, 94(5): 893–904. doi: 10.1111/j.1365-2745.2006.01158.x
- Bormann H, Klaassen K, 2008. Seasonal and land use dependent variability of soil hydraulic and soil hydrological properties of two Northern German soils. *Geoderma*, 145(3–4): 295–302. doi: 10.1016/j.geoderma.2008.03.017
- Boyle M, Frankenberger Jr W T, Stolzy L H, 1989. The influence of organic matter on soil aggregation and water infiltration. *Journal of Production Agriculture*, 2: 290–299. doi: 10.2134/jpa1989.0290
- Brown S, Cotton M, 2011. Changes in soil properties and carbon content following compost application: results of on-farm sampling. *Compost Science & Utilization*, 19(1): 87–96. doi: 10.1080/1065657x.2011.10736983
- Bruand A, Cousin I, 1995. Variation of textural porosity of a clay-loam soil during compaction. *European Journal of Soil Science*, 46(3): 377–385. doi: 10.1111/j.1365-2389.1995.tb01334.x
- Bu Hongmei, Dang Haishang, Zhang Quanfa, 2010. Impacts of forest vegetation on water environment of the Jinshui River in the upper Han River. *Acta Ecologica Sinica*, 30(5): 1341–1348. (in Chinese)
- Celik I, Gunal H, Budak M et al., 2010. Effects of long-term organic and mineral fertilizers on bulk density and penetration resistance in semi-arid Mediterranean soil conditions. *Geoderma*, 160(2): 236–243. doi: 10.1016/j.geoderma.2010.09.028
- Chai Yafan, Wang Enheng, Chen Xiangwei, 2007. Water storage capacity and permeability of undisturbed typical black soil. *Journal of Soil and Water Conservation*, 21(3): 158–161. (in Chinese)
- Chen Shuai, Wang Xiaoke, Lu Fei, 2012. Research on forest microbial community function variations in urban and suburban forests. *Chinese Journal of Soil Science*, 43(3): 614–620. (in Chinese)
- Chen Y J, Day S D, Wick A F et al., 2013. Changes in soil carbon pools and microbial biomass from urban land development and subsequent post-development soil rehabilitation. *Soil Biology and Biochemistry*, 66: 38–44. doi: 10.1016/j.soilbio.2013.06.022
- Chen Y J, Day S D, Wick A F et al., 2014. Influence of urban land development and subsequent soil rehabilitation on soil aggregates, carbon, and hydraulic conductivity. *Science of the Total Environment*, 494–495: 329–336. doi: 10.1016/j.scitotenv.2014.06.099
- Cheng D B, Dong L Y, Qian F et al., 2017. Observation and modeling on irregular purple soil water infiltration process. *Journal of Mountain Science*, 14(6): 1076–1085. doi: 10.1007/s11629-015-3737-x
- Cheng Jiang, Yang Kai, Lü Yongpeng et al., 2009. Experimental study on rainfall-runoff pollutant reduction by urban green space. *Chinese Journal of Environmental Science*, 30(11):

- 3236–3242. (in Chinese)
- Dornauf C, Burghardt W, 2000. The effects of biopores on permeability and storm infiltration—case study of the construction of a school. *First International Conference on Soils of Urban, Industrial, Traffic and Mining Areas*. Essen: University of Essen, 459–164
- Ellis T W, Leguédou S, Hairsine P B et al., 2006. Capture of overland flow by a tree belt on a pastured hillslope in south-eastern Australia. *Soil Research*, 44(2): 117–125. doi: 10.1071/SR05130
- Endreny T, Collins V, 2009. Implications of bioretention basin spatial arrangements on stormwater recharge and groundwater mounding. *Ecological Engineering*, 35(5): 670–677. doi: 10.1016/j.ecoleng.2008.10.017
- Fang Xi, Tang Zhijuan, Tian Dalun et al., 2012. Distribution and ecological risk assessment of 7 heavy metals in urban forest soils in Changsha city. *Acta Ecologica Sinica*, 32(23): 7596–7606. (in Chinese)
- Fischer C, Roscher C, Jensen B et al., 2014. How do earthworms, soil texture and plant composition affect infiltration along an experimental plant diversity gradient in grassland? *PLoS One*, 9(2): e98987. doi: 10.1371/journal.pone.0098987
- Fu X L, Shao M A, Wei X R et al., 2010. Soil organic carbon and total nitrogen as affected by vegetation types in Northern Loess Plateau of China. *Geoderma*, 155(1–2): 31–35. doi: 10.1016/j.geoderma.2009.11.020
- Gao Y, Zhu B, Zhou P et al., 2009. Effects of vegetation cover on phosphorus loss from a hillslope cropland of purple soil under simulated rainfall: a case study in China. *Nutrient Cycling in Agroecosystems*, 85(3): 263–273. doi: 10.1007/s10705-009-9265
- Gill S E, Handley J F, Ennos A R et al., 2007. Adapting cities for climate change: the role of the green infrastructure. *Built Environment*, 33(1): 115–133. doi: 10.2148/benv.33.1.115
- Guo L B, Gifford R M, 2002. Soil carbon stocks and land use change: a meta analysis. *Global Change Biology*, 8(4): 345–360. doi: 10.1046/j.1354-1013.2002.00486.x
- Hamilton G W, Waddington D V, 1999. Infiltration rates on residential lawns in central Pennsylvania. *Journal of Soil and Water Conservation*, 54(3): 564–568.
- He Xingyuan, 2004. *Shenyang Urban Forest*. Beijing: Science Press, 41–44. (in Chinese)
- Hobbie S E, Ogdahl M, Chorover J et al., 2007. Tree species effects on soil organic matter dynamics: the role of soil cation composition. *Ecosystems*, 10(6): 999–1018. doi: 10.1007/s10021-007-9073-4
- Hood M J, Clausen J C, Warner G S, 2007. Comparison of stormwater lag times for low impact and traditional residential development. *Journal of the American Water Resources Association*, 43(4): 1036–1046. doi: 10.1111/j.1752-1688.2007.00085.x
- Hooker T D, Compton J E, 2003. Forest ecosystem carbon and nitrogen accumulation during the first century after agricultural abandonment. *Ecological Applications*, 13(2): 299–313. doi: 10.1890/1051-0761(2003)013[0299:FECANA]2.0.CO;2
- Hou Xiaolong, Zhuang Kai, Liu Aiqin et al., 2012. Restoration of soil quality after mixed-species planting on mining wasteland at Zijinshan gold-copper mine, Fujian Province, China. *Journal of Agro-Environment Science*, 31(8): 1505–1511. (in Chinese)
- Huang D C, Su Z M, Zhang R Z et al., 2010. Degree of urbanization influences the persistence of *Dorytomus* weevils (Coleoptera: Curculionidae) in Beijing, China. *Landscape and Urban Planning*, 96(3): 163–171. doi: 10.1016/j.landurbplan.2010.03.004
- Huang Xin, Huang Xiaojun, Chen Cai, 2009. The characteristic, mechanism and regulation of urban spatial expansion of Changchun. *Areal Research and Development*, 28(5): 68–72. (in Chinese)
- Institute of Soil Science, Chinese Academy of Sciences, 1978. *Soil Physical and Chemical Analysis*. Shanghai: Shanghai Science and Technology Press. (in Chinese)
- Jia H F, Yao H R, Tang Y et al., 2015. LID-BMPs planning for urban runoff control and the case study in China. *Journal of Environmental Management*, 149: 65–76. doi: 10.1016/j.jenvman.2014.10.003
- Jim C Y, 1993. Soil compaction as a constraint to tree growth in tropical and subtropical urban habitats. *Environmental Conservation*, 20(1): 35–49. doi: 10.1017/S037689290003
- Jim C Y, 1998a. Physical and chemical properties of a Hong Kong roadside soil in relation to urban tree growth. *Urban Ecosystems*, 2(2–3): 171–181. doi: 10.1023/A:1009585700191
- Jim C Y, 1998b. Urban soil characteristics and limitations for landscape planting in Hong Kong. *Landscape and Urban Planning*, 40(4): 235–249. doi: 10.1016/S0169-2046(97)00117-5
- Jim C Y, 1998c. Soil characteristics and management in an urban park in Hong Kong. *Environmental Management*, 22(5): 683–695. doi: 10.1007/s002679900139
- Joffre R, Rambal S, 1993. How tree cover influences the water balance of Mediterranean rangelands. *Ecology*, 74(2): 570–582. doi: 10.2307/1939317
- Kang Wenxing, Guo Qinghe, He Jienan et al., 2008. Function and value analysis of water conservation, soil reinforcement and fertility maintenance of urban forest in Guangzhou. *Scientia Silvae Sinicae*, 44(1): 19–25. (in Chinese)
- Kohnke N, 1968. *Soil Physics*. New York: McGraw-Hill.
- Kumar S, Anderson S H, Udawatta R P et al., 2012. Water infiltration influenced by agroforestry and grass buffers for a grazed pasture system. *Agroforestry Systems*, 84(3): 325–335. doi: 10.1007/s10457-011-9474-4
- Li Jianxing, He Binghui, Mei Xuemei et al., 2013. Effects of different planting modes on the soil permeability of sloping farmlands in purple soil area. *Chinese Journal of Applied Ecology*, 24(3): 725–731. (in Chinese)
- Li Jinli, Yue Xiaojing, Sun Kui et al., 2015. Investigation and improvement of soil fertility of road greenbelt in Tianjin city. *Urban Environment & Urban Ecology*, 28(6): 17–21. (in Chinese)
- Li Yiman, Xiu Chunliang, Wei Ye et al., 2012. Analysis on

- mechanism and spatial-temporal features of urban sprawl: a case study of Changchun. *Economic Geography*, 32(5): 59–64. (in Chinese)
- Liu Daoping, Chen Sanxiong, Zhang Jinchi et al., 2007. Soil infiltration characteristics under main vegetation types in Anji County of Zhejiang Province. *Chinese Journal of Applied Ecology*, 18(3): 493–498. (in Chinese)
- Lu Rukun, 2000. *Soil and Agricultural Chemical Analysis Methods*. Beijing: Chinese Agricultural Scientific and Technology Press. (in Chinese)
- Lu S G, Wang H Y, Bai S Q, 2009. Heavy metal contents and magnetic susceptibility of soils along an urban-rural gradient in rapidly growing city of eastern China. *Environmental Monitoring & Assessment*, 155(1–4): 91–101. doi: 10.1007/s10661-008-0420-5
- Lv H L, Wang W J, He X Y et al., 2016. Quantifying tree and soil carbon stocks in a temperate urban forest in Northeast China. *Forests*, 7(9): 200. doi: 10.3390/f7090200
- Martens D A, Frankenberger Jr W T, 1992. Modification of infiltration rates in an organic-amended irrigated. *Agronomy Journal*, 84(4): 707–717. doi: 10.2134/agronj1992.00021962008400040032x
- Meek B D, Rechel E A, Carter L M et al., 1989. Changes in infiltration under alfalfa as influenced by time and wheel traffic. *Soil Science Society of America Journal*, 53(1): 238–241. doi: 10.2136/sssaj1989.03615995005300010042x
- Merilä P, Malmivaara-Lämsä M, Spetz P et al., 2010. Soil organic matter quality as a link between microbial community structure and vegetation composition along a successional gradient in a boreal forest. *Applied Soil Ecology*, 46(2): 259–267. doi: 10.1016/j.apsoil.2010.08.003
- Milesi C, Elvidge C D, Nemani R R et al., 2003. Assessing the impact of urban land development on net primary productivity in the southeastern United States. *Remote Sensing of Environment*, 86(3): 401–410. doi: 10.1016/S0034-4257(03)00081-6
- Nie Fahui, Li Tian, Yao Haifeng, 2008. Characteristics of soil samples of five Shanghai urban green areas and their effects on reduction of storm water runoff. *Environmental Pollution & Control*, 30(2): 49–52. (in Chinese)
- Oldfield E E, Felson A J, Wood S A et al., 2014. Positive effects of afforestation efforts on the health of urban soils. *Forest Ecology and Management*, 313: 266–273. doi: 10.1016/j.foreco.2013.11.027
- Pan Yongjun, Chen Bufeng, Xiao Yihua et al., 2008. Heavy metal pollution status and evaluation of urban forest soils in Guangzhou. *Ecology and Environment*, 17(1): 210–215. (in Chinese)
- Pataki D E, Carreiro M M, Cherrier J et al., 2011. Coupling biogeochemical cycles in urban environments: ecosystem services, green solutions, and misconceptions. *Frontiers in Ecology and the Environment*, 9(1): 27–36. doi: 10.1890/090220
- Pitt R, Chen S E, Clark S E et al., 2008. Compaction's impacts on urban storm-water infiltration. *Journal of Irrigation and Drainage Engineering*, 134(5): 652–658. doi: 10.1061/(ASCE)0733-9437(2008)134:5(652)
- Pitt R, Lantrip J, Henry C L et al., 1999. Infiltration through disturbed urban soils and compost-amended soil effects on runoff quality and quantity. US Environmental Protection Agency, Water Supply and Water Resources Division, National Risk Management Research Laboratory, Cincinnati, Ohio.
- Pouyat R V, McDonnell M J, 1991. Heavy metal accumulations in forest soils along an urban-rural gradient in Southeastern New York, USA. *Water, Air, and Soil Pollution*, 57(1): 797–807. doi: 10.1007/BF00282943
- Pouyat R, Groffman P, Yesilonis I et al., 2002. Soil carbon pools and fluxes in urban ecosystems. *Environmental Pollution*, 116(S1): S107–S118. doi: 10.1016/S0269-7491(01)00263-9
- Pouyat R V, Yesilonis I D, Szlavecz K et al., 2008. Response of forest soil properties to urbanization gradients in three metropolitan areas. *Landscape Ecology*, 23(10): 1187–1203. doi: 10.1007/s10980-008-9288-6
- Radke J K, Berry E C, 1993. Infiltration as a tool for detecting soil changes due to cropping, tillage, and grazing livestock. *American Journal of Alternative Agriculture*, 8(4): 164–174. doi: 10.1017/s0889189300005385
- Regués D, Badía D, Echeverría M T et al., 2017. Analysing the effect of land use and vegetation cover on soil infiltration in three contrasting environments in Northeast Spain. *Geographical Research Letters*, 43(1): 141–169. doi: 10.18172/cig.3164
- Ren Z B, He X Y, Zheng H F et al., 2013. Estimation of the relationship between urban park characteristics and park cool island intensity by remote sensing data and field measurement. *Forests*, 4(4): 868–886. doi: 10.3390/f4040868
- Richard G, Cousin I, Sillon J F et al., 2001. Effect of compaction on the porosity of a silty soil: influence on unsaturated hydraulic properties. *European Journal of Soil Science*, 52(1): 49–58. doi: 10.1046/j.1365-2389.2001.00357.x
- Saito T, Yasuda H, Suganuma H et al., 2016. Predicting soil infiltration and horizon thickness for a large-scale water balance model in an arid environment. *Water*, 8(3): 96. doi: 10.3390/w8030096
- Sanders R A, 1986. Urban vegetation impacts on the hydrology of Dayton, Ohio. *Urban Ecology*, 9(3–4): 361–376. doi: 10.1016/0304-4009(86)90009-4
- Saxton K E, Rawls W J, Romberger J S et al., 1986. Estimating generalized soil-water characteristics from texture. *Soil Science Society of America Journal*, 50(4): 1031–1036. doi: 10.2136/sssaj1986.03615995005000040039x
- Scanlan C A, Hinz C, 2010. Insights into the processes and effects of root-induced changes to soil hydraulic properties. *19th World Congress of Soil Science, Soil Solutions for a Changing World*. Brisbane: DVD, 1–6.
- Shan Qihua, Yu Yuanchun, Zhang Jianfeng et al., 2009. Comprehensive estimation of soil fertility in urban forest. *Bulletin of Soil and Water Conservation*, 29(4): 186–190. (in Chinese)
- Shaw R K, Wilson M A, Reinhardt L et al., 2010. Geochemistry of artificial coarse fragment types from selected New York City soils. In: *19th World Congress of Soil Science. Soil Solutions for a Changing World*. Brisbane, Australia: DVD.
- United Nations, 2012. *World population prospects: The 2011*

- version. New York: United Nations Department of Economic and Social Affairs, Population Division.
- Wang Jinye, Li Haifang, Duan Wenjun et al., 2013. Runoff processes and the influencing factors in a small forested watershed of upper reaches of Lijiang River. *Scientia Silvae Sinicae*, 49(6): 149–153. (in Chinese)
- Wei Junling, Jin Youqian, Gao Hongjian et al., 2012. Investigation on soil water infiltration in different urban green lands in Hefei City. *Chinese Agricultural Science Bulletin*, 28(25): 302–307. (in Chinese)
- Wei Xi, Tang Ningyuan, Li Tian, 2011. Field survey and improvement for soil infiltration characteristics of urban green land in Shanghai. *Water Purification Technology*, 30(4): 78–83. (in Chinese)
- Winzig G, 2000. The concept of storm water infiltration. In: *First International Conference on Soils of Urban, Industrial, Traffic and Mining Areas*. Essen: Essen University Press, 427–433.
- Woltemade C J, 2010. Impact of residential soil disturbance on infiltration rate and stormwater runoff. *Journal of the American Water Resources Association*, 46(4): 700–711. doi: 10.1111/j.1752-1688.2010.00442.x
- Wong F, Harner T, Liu Q T et al., 2004. Using experimental and forest soils to investigate the uptake of polycyclic aromatic hydrocarbons (PAHs) along an urban-rural gradient. *Environmental Pollution*, 129(3): 387–398. doi: 10.1016/j.envpol.2003.12.006
- Wu G L, Yang Z, Cui Z et al., 2016. Mixed artificial grasslands with more roots improved mine soil infiltration capacity. *Journal of Hydrology*, 535: 54–60. doi: 10.1016/j.jhydrol.2016.01.059
- Xi Zhuxiang, Zhang Chenchen, Lv Zhong et al., 2016. The risk pre-estimation of the flood casualty loss caused by heavy rainstorm in Jilin Province. *Journal of Glaciology and Geocryology*, 38(2): 395–401. (in Chinese)
- Xiao Ronghuan, Li Tiecheng, Qiao Na, 1989. Environmental geomorphology and urban construction of Changchun city. *Scientia Geographica Sinica*, 9(3): 221–232. (in Chinese)
- Yang Jinling, Zhang Ganlin, Zhao Yuguo et al., 2005. Application and comparison of soil compaction indexes in the evaluation of urban soils. *Transactions of the Chinese Society of Agricultural Engineering*, 21(5): 51–55. (in Chinese)
- Yang Jinling, Zhang Ganlin, 2008. Loss of soil water capacity in urban areas and its impacts on environment. *Soils*, 40(6): 992–996. (in Chinese)
- Yang Jinling, Zhang Ganlin, Yuan Dadang, 2008. Characteristics of water infiltration in urban soils of Nanjing city. *Chinese Journal of Applied Ecology*, 19(2): 363–368. (in Chinese)
- Yang J L, Zhang G L, 2011. Water infiltration in urban soils and its effects on the quantity and quality of runoff. *Journal of Soils and Sediments*, 11(5): 751–761. doi: 10.1007/s11368-011-0356-1
- Yao L, Chen L D, Wei W et al., 2015. Potential reduction in urban runoff by green spaces in Beijing: a scenario analysis. *Urban Forestry & Urban Greening*, 14(2): 300–308. doi: 10.1016/j.ufug.2015.02.014
- Yimer F, Ledin S, Abdelkadir A, 2006. Soil organic carbon and total nitrogen stocks as affected by topographic aspect and vegetation in the Bale Mountains, Ethiopia. *Geoderma*, 135: 335–344. doi: 10.1016/j.geoderma.2006.01.005
- Yoder R E, 1936. A direct method of aggregate analysis of soils and a study of the physical nature of erosion losses. *Agronomy Journal*, 28(5): 337–351.
- Zhai C, Wang W J, He X Y et al., 2017. Urbanization drives SOC accumulation, its temperature stability and turnover in forests, Northeastern China. *Forests*, 8(4): 130. doi: 10.3390/f8040130
- Zhang B, Xie G D, Zhang C Q et al., 2012. The economic benefits of rainwater-runoff reduction by urban green spaces: a case study in Beijing, China. *Journal of Environmental Management*, 100: 65–71. doi: 10.1016/j.jenvman.2012.01.015
- Zhang B, Xie G D, Li N et al., 2015. Effect of urban green space changes on the role of rainwater runoff reduction in Beijing, China. *Landscape and Urban Planning*, 140: 8–16. doi: 10.1016/j.landurbplan.2015.03.014
- Zhang Dan, Zheng Haifeng, Ren Zhibin et al., 2015b. Effects of forest type and urbanization on carbon storage of urban forests in Changchun, Northeast China. *Chinese Geographical Science*, 25(2): 147–158. doi: 10.1007/s11769-015-0743-4
- Zhang J, Lei T W, Qu L Q et al., 2017. Method to measure soil matrix infiltration in forest soil. *Journal of Hydrology*, 552: 241–248. doi: 10.1016/j.jhydrol.2017.06.032
- Zhang M K, Wang M Q, Liu X M et al., 2003. Characterization of soil quality under vegetable production along an urban-rural gradient. *Pedosphere*, 13(2): 173–180.
- Zhang M K, 2004. Phosphorus accumulation in soils along an urban-rural land use gradient in Hangzhou, Southeast China. *Communications in Soil Science and Plant Analysis*, 35(5–6): 819–833. doi:10.1081/css-120030360
- Zhao Y G, Zhang G L, Zepp H et al., 2007. Establishing a spatial grouping base for surface soil properties along urban-rural gradient—A case study in Nanjing, China. *Catena*, 69(1): 74–81. doi: 10.1016/j.catena.2006.04.017
- Zhao Z X, Guo H C, 2010. Effects of urbanization on the quantity changes of microbes in urban-to-rural gradient forest soil. *Agricultural Science & Technology*, 11(3): 118–122.